

MINISTRY OF HEALTH

Assessment of blood lead levels among children in Owino Ouru Settlement in Mombasa County, Kenya, 2015

By Lead Poisoning Investigation Team

May 2015

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Executive Summary

Background: Concerns of possible lead exposure from fly ash emissions of a battery recycling factory in an informal settlement in Mombasa was brought up by the community members following high lead levels found in blood samples from three children. This study aimed at determining blood lead concentrations, lead levels in the environment and factors associated with elevated blood lead level among children aged 12 to 59 months from the settlement.

Methods: This is a cross sectional study of children aged 12-59 randomly selected from households in two informal neighbouring settlements (Owino Ouru and Bangladesh). Structured questionnaires on social, demographic, child's behaviour and household characteristics were administered to caregivers. Venous blood (1-3ml) drawn from each child was tested for lead using LeadCare®II portable analyzer. Household dust, drinking water and soil from compound in half of the sampled households was collected and tested for lead using graphite furnace atomic spectrometry at the Government chemist.

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Results: We obtained blood samples from 161 children in 161 households, 83 soil samples, 76 dust and 73 water samples. Of these, 31 blood samples were not included in the analysis because they were obtained during pretesting and upon parents' request. We randomly selected 130 children, 65 from each settlement, of these 59 (45%) were males and median age was 39 months (Interquartile Range (IQR): 30 - 52). Blood lead levels ranged from 1µg/dL to 31µg/dL with 20 (31%) children from Owino Ouru and 5 (8%) children from Bangladesh having blood lead levels of ≥ 10µg/dL. Forty five (69%) children from Owino Ouru and 18 (28%) children from Bangladesh had blood lead levels of ≥ 5µg/dL. None of the children had blood lead levels above 45µg/dL. The mean blood lead level (9µg/dL; Standard deviation (SD): 6) of children from Owino Ouru was significantly higher than the mean blood lead level (4.4 μ g/dL; SD: 3) of children from Bangladesh, t= 5.47; 95% CI: 2.9 - 6.2; (p = < 0.0001). Nine (30%) of environmental samples from Owino Ouru and 15 (47%) of those from Bangladesh had lead concentrations in either soil >400mg/Kg, dust >40µg/f² or drinking water (>10µg/L or 0.01mg/L). The mean lead concentration of soil (387 mg/Kg; SD: 625) from Owino Ouru was also significantly higher than the mean lead concentration of soil (74 mg/Kg; SD: 127) from Bangladesh, t= 2.77; 95% CI: 87 – 539; (p = 0.007). The presence of soil lead > 400 mg/Kg was associated with higher blood lead levels (p=0.01).

Conclusions: There were significantly high blood lead levels in children from Owino Ouru indicating that the children are being exposed to lead in their living environment. This requires medical care, blood lead screening, community sensitization, health education, nutritional counselling, surveillance and environmental lead reduction interventions.

Key words: Childhood, Blood, Lead, Environmental Exposure, Kenya

Introduction

It is estimated that high blood lead levels contribute to approximately 600,000 cases of intellectual disability in children annually (WHO, 2010) and was responsible for 143,000 deaths in 2004. There is no known safe threshold of blood exposure since adverse effects have been shown to occur at levels of below 5µg/dL (1). Low lead levels in young children results in mental retardation, impaired intellectual and cognitive function, learning disabilities, poor attention, hearing disorders and decreased growth (2). Acute lead poisoning causes gastrointestinal disturbances (anorexia, nausea, vomiting, abdominal pain), hepatic and renal damage and neurological effects (malaise, drowsiness, encephalopathy) that may lead to convulsions, coma and death (3). Lead exposure is often asymptomatic at low levels hence remains undiagnosed and untreated.

Children <5 years old are at increased risk of lead poisoning because of their innate curiosity and hand to mouth behaviour (4). Children have higher lead intake per unit body weight and body systems not fully developed leading to increased gastrointestinal absorption and distribution to the brain through the blood brain barrier (5). Their developing central nervous system is also more vulnerable to lead. Absorption of lead into the body is affected by many factors, including age, nutritional status and lead particulate size. Lead is poorly excreted, and most lead is sequestered in bone. As a result, elevated BLLs take months to years to decrease, even in cases where external exposures have been well controlled and chelation therapy has been instituted, because bone stores are mobilized into blood (6).

During the last decade measures have been taken to reduce lead exposure globally. Following the removal of lead from gasoline, dramatic reductions in blood lead concentrations have been observed (7) (8) however, leaded paint in houses built before 1950 remains the major source of lead exposure for American children (9). Children are exposed to lead through ingestion of weathering paint and inhaling dust or soil contaminated with such paint. In low income countries children are at increased risk of exposure to lead from other multiple sources (10). Studies have documented elevated blood lead levels in children living in communities involved in mining (11) smelting (12) and used acid lead battery recycling (13) where the children get exposed to lead by eating the contaminated soil or inhaling polluted air from the industrial emissions. Additionally, outbreaks of lead intoxication have been associated with drinking water from pipes, solders and fittings containing lead, use of ceramic utensils glazed with lead, eating food stored in lead soldered cans, playing with toys

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containing lead, use of traditional remedies such as *kohl* and cosmetics containing lead (10). Adults working in jobs or hobbies such as smelting, auto repair, firing ranges, painting, ceramics, pottery, electrical, wire and cable works, battery manufacturing and recycling and stained glass making, carry the generated lead dust on clothes and contribute to the lead dust in the child's home. Children living in these homes are exposed to lead by licking dust laden fingers or inhaling lead particles in the air.

In 2007 a battery recycling and smelting factory was established in an informal settlement in Mombasa County along coastal Kenya. In 2010 an investigation was conducted in the Owino Ouru and three children were tested for blood lead levels. They had blood lead levels as high as 23µg/dl, 17µg/dl and 12µg/dl (Okeyo, 2012). Following these findings community members complained and the factory was closed in July 2012 by the local health authorities. Despite this action, concerns remained regarding appropriate closure of the factory and possible ongoing lead exposures in the community.

On 21st July 2014, the Ministry of Health (MOH) began an investigation to determine blood lead levels among children aged 12 to 59 months in the informal settlement, lead levels in the environment and identify risk factors for elevated blood lead level in this population. Findings from the investigation will enable public health authorities to make evidence based actions to safeguard public health in the affected community.

Methods

Site

We conducted the investigation in two informal settlements located in Mombasa County along coastal Kenya (Figure 1). Owino Ouru borders the northern side of the battery recycling factory, and consists of 450 households with a population of 1,700 persons. Bangladesh, a community located on the south western side and more than 2km upwind from the factory was included as the comparison site. Bangladesh consists of 1,500 households with a population of 5,000 persons. Bangladesh has similar demographics (such as traffic patterns, housing conditions) but does not share water sources with Owino Ouru and has no known battery recycling factories.

Investigation

We enrolled children aged 12 to 59 months who resided in either of the settlements since January 2014. We investigated children because at this age they are vulnerable to relatively low lead levels and develop irreversible health effects. We excluded children from the survey whose parents or guardians did not provide consent and households that were vacant at time of household listing. Village elders from the two settlements filled in a household list with details of household membership and ensured that all eligible households with children 12-59 months of age were included in the sampling frame.

When we assumed blood lead level variance in this population to be 25 and 10% non-response rate, a sample size of 65 children from each settlement provided us 99% power to detect a difference in mean blood lead levels of 4µg/dL at 95% confidence. We randomly selected 65 households from each sampling frame using table of random numbers. Upon reaching a house we administered four screening questions to whoever was found in the house. These questions included whether the household has a child aged 12 − 59 months, duration of stay in the settlement ≥12 months, presence of parent or guardian and whether they will provide a written consent. If the answers to all the four questions were 'yes' we proceeded with the interview. One child was randomly selected to participate in the study. If a house was unoccupied or the sampled child was not present at time of visit, the house was visited later that day or at different time on another day. If a house was permanently vacant, if the caregiver declined to participate or was not available for interview after multiple attempts

then the next household in the sampling frame was visited. We recorded this information on a household visitation log sheet.

A standard explanation describing the purpose of the study, the procedures to be followed, the risks and benefits of participation was read to all caregivers and a signed written consent obtained. Data was collected using a pretested structured questionnaire translated to Kiswahili. We collected demographic information including sex and age, clinical and treatment history, factors known to increase risk for elevated blood lead levels such as children's behaviour habits (frequency of outdoor playing, washing hands before eating, eating soil or sand), housing type, sources of drinking or cooking water, the presence of a household member with known exposure to lead, proximity to a battery recycling factory, and whether waste from factory stand around the house. Global positioning system (GPS) coordinates of place of residence was taken at the entrance of the house and used to map and calculate distances from the smelter factory.

Laboratory

A venous blood specimen was collected from every child. The venepuncture site was thoroughly cleaned with alcohol wipes, dried with dry gauze before the specimen was obtained. Approximately 1 -3ml of blood sample was collected in a vacutainer containing an anticoagulant, ethylenediamine tetra acetic acid (EDTA). In 10% of the children a second sample was obtained for quality control. Lead free blood collection supplies were provided by the Inorganic Chemistry Laboratory at US CDC. All blood samples were stored at 4°C and venous samples were shipped on frozen packs to be analyzed at the National Public Health Laboratory in Nairobi accompanied by laboratory request form.

The blood samples were tested for lead using LeadCare ® II, a portable blood lead analyzer. The LeadCare ® II instrument quantifies BLLs from 3.3- 65 μ g/dL. These levels are measured with a level of detection (LOD) accuracy level of \pm 3 μ g/dL. We used the actual value of levels below the LOD for statistical analyses. However, parents and clinical health care providers were only notified that the child's BLL is below the LOD.

Environmental Sampling

We collected dust, water and soil samples from fifty percent of randomly visited households as per the US Housing and Urban Development (HUD) protocol sampling procedures

recognized by EPA (HUD, 1996). We collected five composite soil samples from bare soil outside the households where the child normally plays. A subsample was obtained from the center and one subsample from each of four different directions one meter from the center. The subsamples were collected by scooping the top surface soil using a four oz plastic scoop and poured into a labelled zip-lock bag. We collected water used by the household for drinking into a 125ml sample bottle. Dust was collected from the floor of the house by multidirectional wipe sampling method using wet wipes over an area laid using reusable template measuring 6 by 6 inches. The folded wipe was placed in a zip lock bag. All the used materials were discarded in trash bags. To prevent contamination or exporting of lead dust, hands were cleaned with baby wipes after each sample collection and clean gloves worn before collecting samples in the next area. All environmental samples were placed separately in properly labeled clean, plastic bags. Sample forms were filled with details that included unique sample number, gprs location, sampling method used, date of collection, and the name of the person who collected each sample. The samples together with laboratory form were sent to Government chemist laboratory for analyses. The environmental samples were analyzed for lead levels using graphite furnace atomic absorption spectrometry at the Government chemist laboratory.

Data management and analysis

All data was entered and validated in database using Epi InfoTM 7.1.4 (CDC, Atlanta, GA, USA). We performed descriptive analysis and used the two-sample t-test in the continuous variables. An elevated blood lead level was defined as blood lead concentration $\geq 10 \mu g/dL$ based on WHO guidelines. Contaminated environment was defined as lead concentration in either soil $\geq 400 \text{mg/Kg}$, in house dust $\geq 40 \mu g/f^2$ or in drinking water $\geq 50 \mu g/L$ (0.05 mg/L). The results were presented in tables and figures.

Ethical Considerations

Written informed consent to administer the questionnaire and draw blood was obtained from the caregivers. All the questionnaires, laboratory specimens and other records were identified by unique identity number to maintain participant confidentiality. We obtained additional blood samples from other children in the sampled household when the parent requested but their data was not included in this analysis. All the children who had elevated blood lead levels ($\geq 10 \mu g/dL$) were linked to follow up care as per the CDC guidelines of managing elevated BLL in children (ACCLPP, 2012). Questionnaires were stored under lock and key.

Electronic data was stored in a password protected database. The scientific approval for the study was obtained from the Ministry of Health, lead poisoning Technical Working Group (TWG). Expedited ethical approval was sought from the Kenya Medical Research Institute (KEMRI) since this investigation was considered a public health response. Prior to commencement of the investigation, permission was sought from the Mombasa County Department of Health Services.

Results

We visited 161 households and obtained blood samples from 161 children in both settlements. Responses from 31 participants were excluded from data analysis because data and blood samples were obtained during pretesting or from the children whose parents requested but were not in the sampling list. A total of 130 children aged 12 – 59 months: 65 from each settlement were included in the analysis. Median age was 39 months (IQR: 30 - 52), and 59 (45%) males. The median number of people living in households of sampled children was five (range 2 -13) and the median number of children 12 – 59 months in each sampled household was one child (range: 1 – 4). Of the 65 children from Owino Ouru, 31 (48%) lived within 200m radii from factory and 16 (25%) lived in a household with at least one adult working in any job involving lead for last five years. Nine (14%) of the children from Bangladesh were also living in a household with at least one adult working in any job involving lead for last 5 years (Table 1).

Blood lead levels among sampled children ranged from $1\mu g/dL$ to $31\mu g/dL$. In overall, 63(48%) children had blood lead levels of $\geq 5\mu g/dL$ and 20 (19%) children had blood lead levels of $\geq 10\mu g/dL$. There were 45 (69%) children from Owino Ouru and 18 (28%) children from Bangladesh with blood lead levels of $\geq 5\mu g/dL$. Twenty (31%) children from Owino Ouru and five (8%) children from Bangladesh had blood lead levels of $\geq 10\mu g/dL$. None of the children had blood lead levels above $45\mu g/dL$. The mean blood lead level (9 $\mu g/dL$; Standard deviation (SD): 6) of children from Owino Ouru was significantly higher than the mean blood lead level (4.4 $\mu g/dL$; SD: 3) of children from Bangladesh, t= 5.47; 95% CI: 2.9 $\frac{1}{2}$ 6.2; (p = < 0.0001).

We collected 83 soil samples, 76 dust and 73 water samples. Among these, 62 of each sample type: 30 from Owino and 32 from Bangladesh were included in the analysis. The remainder were obtained from households that recorded high blood lead in children and from the factory and its environs. The mean lead concentration of soil from Owino Ouru (387 mg/Kg; SD: 625) was significantly higher than the mean lead concentration of soil from Bangladesh (74mg/Kg; SD: 127); t=2.77; 95% CI: 87-539; (p=0.007). Soil samples obtained in the factory compound had the highest lead concentration (26,837 mg/Kg) followed by the one collected at the factory gate (2,381 mg/Kg). The mean lead loading of house dust from Owino Ouru (15.6µg/f²; SD: 29) was higher than the mean lead loading of house dust samples from Bangladesh (4.7µg/f²; SD: 14) but this difference was not statistically

significant t= 1.93; 95% CI: 0.4 - 22; (p = 0.058) and both levels were below the action levels. The lead concentration of water samples from Owino Ouru was below the action level ($50\mu g/L$).

In the analysis of children from Owino Ouru only, more males 24 (60%) as compared to females had blood lead levels $\geq 10\mu g/dL$ (OR: 5 (95% CI:2-15). Of the 20 children with blood lead level $\geq 10\mu g/dL$, 19(90%) reside <200m radii from factory. There was no significant difference in blood lead among children living in households that had adults who worked for the last 5 years in jobs that involved lead. The main source of drinking or cooking water in 100% the households was from tanks supplied by the municipality. Children from houses that used other sources of water such as wells 8(40%), rain water 2(10%) and boreholes in 6(30%) did not show significant differences in the blood lead levels. Mud floor 5(25%), mud wall 15(60%), growing own or obtaining vegetables grown by neighbour for consumption 1(5%), eating soil/sand 13(65%), not washing hands all times before eating 9(45%), ever breastfed 65(100%), playing outside >6hrs 13(65%) and spending most time in school 3(15%), in compound 2(10%) or in the house 2(10%) also did not reveal any differences in blood lead (Table 2).

Discussion

Findings from our study demonstrate significant elevations in blood lead levels among children aged 12 – 59 months in Owino Ouru in comparison to similar children from Bangladesh. There were also significant elevations in soil lead concentration in Owino Ouru. Lead content in soil, close proximity of the household to the smelter and sex of child were identified as risk factors that expose children living in Owino Ouru to higher blood lead levels. Our findings support the hypothesis that blood lead levels among children in Owino Ouru are different from the blood lead levels of children from Bangladesh and the difference is associated with increased lead concentrations in the Owino Ouru children's environment. Urgent interventions are required to reduce lead exposure in the affected community.

Although none of the children had blood lead level 45µg/dL or above, the level at which CDC recommends chelating therapy, the resulting study showed that 70% of the tested children in Owino Ouru had levels above the reference value of 5µg/dL and 30% had blood lead levels at or above the WHO maximum acceptable level ≥10µg/dL, a rate 4 times higher than that found in children in a demographically similar community of Bangladesh that does not have a battery factory. This prevalence is higher than the prevalence of 1.4% reported in bio-monitoring studies of U.S. children 1-5 years old in NHANES 1999-2004 (7). Blood lead elevations have similarly been documented in other studies of children exposed to industrial emissions from used acid battery recycling informal establishments. Fifty nine percent of children in Jamaica (12), 80% of Nicaragua children (14) and 91% of children from Haina, Dominican (12) had blood lead level >10µg/dL. Three of the children in Owino Ouru had levels of between 20 and 45µg/dL. At those levels, there could be clinical evidence of lead poisoning (interference with vitamin D metabolism), but the health consequences of the prevailing elevations of blood lead are usually demonstrable only on a population basis as cognitive and developmental deficits. Further studies on neuropsychological and developmental evaluations are necessary. The public health response to such findings should be termination of toxic emissions and abatement of existing contamination; removal of children from the identified sources is usually not an option.

The finding that 8% of the children from Bangladesh had blood lead level >10µg/dL indicates that sources of lead exposure other than battery activities in these communities exist. This finding is however lower than the prevalence of 14% reported among 1-3 year old well Lebanese children presenting to paediatric clinics (15) and 30% of the controls in the

Nicaraguan children (14). Higher lead concentration observed in water samples could be a possible reason for this elevation. There is need to screen for blood lead, monitor water quality and establish the other common causes of lead exposure among the children not necessarily living near battery factory.

Our study had two limitations. First, the study sample size focused on determining difference in blood lead levels in the two settlements. We did not have enough statistical power to study the potential risk factors between the children with different blood lead levels. We also had few adults working in jobs known to involve lead and therefore we could not determine effect of take home lead. Secondly because this cross sectional investigation carried out two years after closure of factory, the magnitude of poisoning may have been different among the children who were in this age category two years ago. The difference in blood lead levels obtained in this study cannot be generalised to that of children living in the settlement when the factory was operational. Despite these limitations it appears that significant childhood lead exposure from the environment still persists, however follow-up assessments are needed to identify all potential sources.

Conclusions

There was significantly high blood lead levels in children from Owino Ouru associated with high soil lead concentrations. The most immediate priority is to reduce exposure to lead and other contaminants. This is best accomplished by developing and implementing a comprehensive and integrated intervention plan. Specifically, the process should ensure that the factory remains closed so as to reduce air lead emissions; Implement interventions that have been demonstrated scientifically to reduce lead exposure from historical soil contamination and develop a scientifically robust plan to monitor the impact of lead reduction efforts. To strengthen the overall process and plan, and to improve credibility and ensure that monitoring and other needs of affected parties are met, stakeholders should participate in reduction planning, implementation, and monitoring of lead and other contaminants.

Laboratory results should be made available to study participants immediately for appropriate action. In 2012 CDC adopted an upper reference range value for blood lead of $5\mu g/dL$ based on the 97.5% of the distribution of BLLs in children. We recommend that children with blood lead levels $\geq 5\mu g/dL$ have venous blood lead tests performed three months after this investigation and follow up of the children over time until the environmental investigations

and subsequent responses are complete. Children with blood lead levels $< 5 \mu g/dL$ require retesting in not less than one year until the lead sources have been controlled or eliminated or they turn 6 years old. In this community irreversible adverse effects of lead poisoning will be prevented if children undergo monitoring of blood lead levels.

Certain vitamins and minerals, especially Calcium, Iron, Zinc and vitamin C, play a specific role in minimizing lead absorption. A well-balanced diet is essential to meeting the child's recommended daily allowance of essential vitamins and minerals and to provide adequate calories for growth. Regular assessment of the child's nutritional status during follow up care can identify children with inadequate intake of these and other nutrients, and allow the clinician to proactively recommend supplementation. Children exposed to lead should be evaluated for anaemia and any iron deficiency corrected using iron supplements (AAP gidelines).

Training of clinicians on high index of suspicion, lead screening questions, blood testing of lead, and treatment is essential in management of lead exposure. Clinicians should emphasize on healthy nutrition and/or dietary supplements. They should also be involved in outreach and community health education, overseeing ongoing monitoring of children with elevated blood lead levels, defined as levels above the reference value, coordinating efforts with parents and county authorities to minimize risks to individual children and to assist communities in their primary prevention efforts.

Primary prevention is a strategy that emphasizes the prevention of lead exposure, rather than a response to exposure after it has taken place. Primary prevention is necessary because the effects of lead appear to be irreversible. In the U.S., this strategy largely requires that children not live in older housing with lead-based paint hazards. Screening children for elevated BLLs and dealing with their housing only when their BLL is already elevated is no longer advisable (Centers for Disease Control and Prevention (CDC) Advisory Committee on Childhood Lead Poisoning Prevention (ACCLP, 2012). The goal of primary prevention is to ensure that all homes become lead-safe and do not contribute to childhood lead exposure. Prevention requires that we reduce environmental exposures before children are exposed to these hazards. It is important to carry out regular environmental testing of dust, soil and water testing and evaluation of homes for lead sources. Termination of toxic emissions and abatement of existing contamination is necessary to reduce exposures. This may require the

removal of children from the identified sources until the environment is confirmed to be safe for the children.

There is need for coordination of care with the local authorities and organizations to plan a response strategy. Efforts to increase awareness of lead hazards and ameliorative nutritional interventions are also key components of a successful prevention policy. Finally the establishment of a screening program and effective screening policies and practices by the Kenya Ministry of Health will ensure that the children of high-risk families are screened, and that lead-exposed children or children with elevated blood lead levels receive key environmental interventions and case management services.

Acknowledgements

This investigation was financially supported by the Ministry of Health, Kenya Field Epidemiology and laboratory training program and testing materials from CDC Atlanta. The authors thank the community members and village elders for their cooperation during the investigation; the assistant Deputy Commissioner, Gitonga and 2 Chiefs for social mobilization; Mombasa County Department for Health services: B.Omar, K. Shikely, R.Mwanyamawi, H.Mohammed, T.Suleiman for permitting us to carry out the investigation; Jomvu Sub-County Health Officers: SCMOH, N.Tindell, H.Wasike, F.Kenyatta, Wycliff; Human Rights activist: P.Omide for mobilizing the community; FELTP staff: A.Sitati, M.Mwangi, Gabriel for administrative support; FELTP Cohort 11 residents for data and sample collection (T.Kigen, Mark, Angeline, V.Oramisi, C.Kiama, Boniface, Maza, Miheso, J.Rotich, Muiruri, Morris, Abdulkadir, Paul, Ngere, B. Ochieng, Caren, Tabitha, Omesa, Githuka); MOH staff: J.Kioko, K.Ombacho; Lead Poisoning Task Force members and MOH staff: H.Onyangore, Shem, P.Ngari, Bunyasi, E.Muniu for scientific review and approval; KEMRI Ethics and Review Committee for ethical approval; Government Chemist staff: B.Wandera, Munyoki, Musyoki, Mombasa staff for laboratory testing; CDC staff Timothy Dignam, Nielsen Jay for developing the study maps and study design; Alfred Musekiwa for providing statistical support and Dorothy L Southern for providing guidance in scientific writing and critically reviewing this report.

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Tables and figures

Table 1: Demographic, clinical and environmental characteristics of children in Lead study in two informal Settlements, Mombasa County, Kenya, January 2015

	Owino Ouru	Bangladesh	Total
	(N = 65)	(N = 65)	(N = 130)
Characteristics	n (%)	n (%)	n (%)
Male	13 (52)	58 (55)	71 (55)
Median age in months (IQR)	42(31-50)	36(27-52)	39(30-52)
Age Groups (Months)			
12 – 23	10 (15)	14 (22)	24 (19)
24 – 35	15 (23)	18 (28)	33 (25)
36 – 47	18 (28)	6 (9)	24 (19)
48 – 59	22 (34)	27 (42)	49 (38)
Living < 200m from Battery recycling			
Factory	31 (48)	0 0)	31 (24)
Living in HH* with adults working in job			
involving lead for last 5 yrs	16 (25)	9 (14)	25 (19)
Γ'ood lead Levels (μg/dL)			
< 5.0	20 (31)	47 (72)	67 (52)
5 - 9.9	25 (39)	13 (20)	38 (29)
10 – 19.9	16 (25)	5 (8)	21 (16)
20 - 45.0	4 (6)	0 (0)	4 (3)
Clinical Symptoms			
Recurrent abdominal Pain	20 (31)	18 (28)	38 (29)
Speech/language delay	11 (17)	2 (3)	13 (10)
Recurrent constipation	6 (9)	4 (6)	10 (8)
Joint pains	4 (6)	2 (3)	6 (5)
Poor concentration	2 (3)	2 (3)	4 (3)
Muscle aches	2 (3)	1 (2)	3 (2)
Seizures	0 (0)	1 (2)	1 (1)
Lead content of environmental samples			
(n)	30	32	62
Soil (>400mg/Kg)	6(20)	1(3)	7(11)
House dust $(>40 \mu g/f^2)$	4(13)	3(9)	7(11)
Drinking water (>10µg/L or 0			
	1(3)	14(44)	15(24)

^{*} Households

Table 2: Analysis of child's behaviour, household characteristics and environment by blood lead level among children aged 12 – 59 months in an informal Settlement, Mombasa County, Kenya, January 2015

Blood Lead Level (μg/dL)					
	>10 (N=20)	< 10 (N=45)		p- Value	
	n (%)	n (%)		•	
Male	24 (60)	12 (53)	5 (2 -15)	0.002	
Age Groups (Months)				0.002	
12 - 23	3 (15)	7 (16)	1 (0.2 - 4)	0.950	
24 - 35	7 (35)	8 (18)	3 (1 - 8)	0.131	
36 – 47	4 (20)	14 (31)	1 (0.2 - 2)	0.359	
48 – 59	6 (30)	16 (36)	1 (0.2 - 2)	0.665	
Living < 200m from Battery	(=)	10 (30)	1 (0.2 - 2)	0.003	
Recycling factory	18 (90)	13 (29)	22 (5 - 109)	< 0.0001	
Adult worked in job that	10 (20)	13 (23)	22 (3 - 109)	\ 0.0001	
involved lead	6 (30)	10 (22.2)	2(1 5)	0.5	
Households with adults	0 (30)	10 (22.2)	_2 (1 - 5)	0.5	
working in following jobs					
for ast 5 yrs.					
Smelting	2 (10)	2 (4)	2 (0 2 10)		
Auto Repair	2 (10)	2 (4)	2 (0.3 - 18)	0.393	
	1 (5)	5 (11)	1 (0.1 - 4)	0.436	
Firing Ranges	0 (0)	2 (4)			
Painting	2 (10)	6 (13)	1 (0.1 - 4)	0.708	
Ceramics	2 (10)	2 (4)	2 (0.3 -18)	1.393	
Electrical	1 (5)	2 (4)	1 (0.1 -14)	0.922	
Battery Manufacturing		0 (0)			
Recycling Batteries	5 (25)	0 (0)			
Wire and Cable	1 (5)	0 (0)			
Pottery	2 (10)	1 (2)	5 (0.4 - 57)	0.171	
Stained glass making	2 (10)	2 (4)	2 (0.3 - 18)	1.393	
Primary Drinking & Cooking					
Water Source					
Public Water	65 (100)	65 (100)			
Well	8 (40)	11 (24)	2 (1 - 6)	0.207	
Rain Water	2(10)	7 (16)	1 (0.1 - 3)	0.207	
Bore hole	6 (30)	10 (22)	2(1-5)	0.505	
Water Vendors	0 (0)	2 (4)	-(1 3)	0.303	
loor Type	(-)	- (.)			
Mud	5 (25)	17 (38)	1 (0.2 - 2)	0.210	
Concrete	13 (65)	26 (58)		0.318	
Vall type	.5 (05)	20 (30)	1 (1 - 4)	0.586	
Mud	15 (60)	76 (72)	1 (0.2 1)	0.000	
Bricks/Stone		76 (72)	1 (0.2 – 1)	0.227	
ources of Vegetables/Fruits	15 (24)	18 (27)	1 (0.4 – 2)	0.690	
onsumed in household					
Grow Own	1 (5)	10 (07)	0.1.(0.00		
	1 (5)	12 (27)	0.1 (0.02 - 1)	0.04	
Neighbour grows	0 (0)	4 (9)			
Vendors	15 (75)	35 (78)	1 (0.3 - 3)	0.808	
ypes of Vegetables consumed					
at are grown by household or					
om neighbour who grows					
Tomatoes	0 (0)	8 (18)			
Kales	0(0)	11 (24)			

Onions Traditional Vegetables Carrots	0 (0) 1 (5) 0 (0)	9 (20) 17 (38) 6 (13)	0.1 (0.01-1)	0.007
Spends most of day In School Inside compound In the house	3 (15) 2 (10) 2 (10)	9 (20) 4 (9)	3 (1 -13) 2 (0.2 - 7)	0.245 0.887
Eat Soil or Sand Doesn't wash hands all times	13 (65) 9 (45)	4 (9) 23 (51) 15 (33)	2 (0.2 - 7) 2 (1 - 5) 2 (1 - 5)	0.887 0.302 0.372

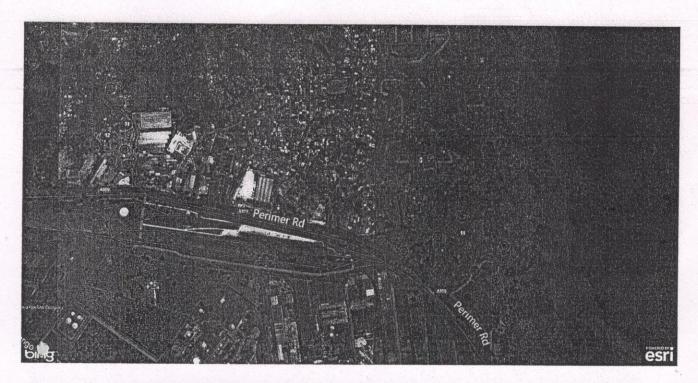


Figure 1: Map showing distribution of children aged 12-59 months in two informal settlements